

Contribuții Botanice – 2011, XLVI: 43-56
Grădina Botanică “Alexandru Borza”
Cluj-Napoca

THE AESTHETICS OF SPONTANEOUS URBAN-INDUSTRIAL VEGETATION

Rüdiger WITTIG

Chair of Ecology and Geobotany, Institute of Ecology, Evolution and Diversity, Max-von-Laue-Straße 13,
D-60438 Frankfurt am Main, Germany
e-mail: r.wittig@bio.uni-frankfurt.de

Abstract: In shrinking cities not all fallow areas can be recycled for urban-industrial use. However, due to financial problems, the remaining open spaces also cannot all be converted into traditional urban parks, because such green spaces require intensive care: they need to be designed, prepared (soil and relief preparation), planted and managed (irrigation, mowing, cutting, replanting). For this reason, in some regions of Europe, particularly in the Ruhr area (Germany), urban-industrial fallows spontaneously covered by vegetation have officially been declared urban green spaces. However, although the arguments are evident, the concept of spontaneous vegetation as larger components within urban green spaces is still not commonly accepted in Central Europe and even more rarely considered outside Europe. The reason is that ruderal vegetation is regarded by most people as an indicator of neglect and abandonment and therefore is seen as ugly and dirty. By considering all available published relevés, this paper shows that a particular group of ruderal communities, the associations of *Dauco-Melilotion*, is species-rich and contains many beautiful eye-catchers as well. Uncoloured flowers (“grey”) never represent more than 30% of the community, mostly even less than 25%. Yellow is most important in all cases, mostly with white on the second and pink-violet in the third position. Blue and yellowish-green are always present, too. Additionally, this vegetation type met all other criteria mentioned in the literature as important for the acceptance of green spaces. It presents its rich colour optimum from June to September, i.e. a period of four months. This is much longer than what can be stated for most flowerbeds in traditional parks. As in large parts of Central Europe summers will become drier and hotter and many of the typical ornamental plants as well as most lawns already needing irrigation under present climate conditions, park planners and managers have to reconsider their concepts with regard to the choice of adapted species. Spontaneous ruderal vegetation is adapted in any case, because its composition is the result of the habitat conditions. Giving more space to this vegetation type would allow more nature experience for city dwellers and more leisure activities than traditional green (because it is more tolerant of disturbance), contribute to biodiversity, save costs and water, and thus represent a substantial step towards a sustainable city.

Key words: Brown field, *Dauco-Melilotion*, ruderal vegetation, spontaneous vegetation, shrinking cities, sustainable city, urban fallows, urban green spaces

Introduction

Numerous papers have been published on the value of urban nature, [e.g. 1, 7, 83, 44, 79]. “Promotion of the development of spontaneous nature in the inner city” is one of the demands that have to be fulfilled by an “ideal city from an ecological view” [86]. Sometimes „aesthetics“ is mentioned as one of these values; however, in such cases generally remnants of natural biotopes [e.g. 8] or urban forests, parks and public gardens are implied (see the reviews of [72, 67]). Galindo & Rodriguez [16] show that „naturalness“ and, within this main feature, the sub-feature „existence of vegetation“, are the most important criteria for giving the attribute “highly aesthetic” to an urban site. In contrast, the same authors identify “lack of maintenance“ and, in particular, „neglect/abandonment“, as the second important reason (behind „lack of naturalness“, in particular “lack of vegetation”) for giving low aesthetic values. Ruderal vegetation is regarded by most people as an indicator of just this unappreciated status of neglect and abandonment. Case studies in Germany asking about the attractiveness of different types of urban green spaces revealed a lower rank for wasteland than for the municipal park, the botanic garden and green courtyards; however, an equal rank as that of highly structured vegetation as can be found in some district parks and in green squares [42]. In group discussions on the value

of protection of different types of green spaces, green fallow land of former industrial areas received the lowest rank [60]. Even to people that use such areas for recreation and leisure activities, spontaneous urban vegetation may be nevertheless associated with mess and noise [28].

Traditional urban green spaces, however, require an intensive amount of care: instead of occurring spontaneously and structuring themselves they have to be designed in advance, constructed (soil preparation) and built. Trees, shrubs and herbs have to be planted and grass needs to be sown. These initial measures are followed by the regular demands of maintenance: cutting of trees and shrubs, fertilizing of flowers, mowing of lawns, and, as necessary, watering during periods of drought.

Especially, popular and therefore often planted, abundantly blossoming ornamental plants seldom meet their natural requirements. Therefore, they react very sensitively to disturbances that cause them to grow weakly and/or die back – thus creating the need to plant anew. Already Hard [20] has indicated that cultivation and conservation of spontaneous vegetation is more reasonable from an ecological and economical point of view, and, moreover, the utilization of spontaneous vegetation allows a broader range of activities. This is demonstrated with special regard to children's playgrounds [23]. Although the arguments are evident, the concept of spontaneous vegetation as larger components within urban green spaces is still not commonly accepted in Central Europe.

Only former areas of coal and steel industry form an exception to the rule. After the closing down of the respective industries, the spacious industrial areas and the accompanying transportation areas suddenly turned into large fallows. Thus neither time nor financial capacities allowed their transformation into traditional urban parks. Some of these German industrial and traffic fallows were declared as nature preserves or places for nature discovery [58]. Thoughts about conservation on „brown fields“ in urban areas also circulate widely in the UK [e.g. 17]. In America [e.g. 35] and Asia [e.g. 31], papers on „urban green space“ almost exclusively deal with parks and remnants of the natural or agricultural landscape.

With the exception of extreme habitats, such as regularly flooded areas, in the long run in Central Europe all vegetation turns into woodlands. In the Ruhr district [75] and Berlin [39], many urban fallows have already turned into urban woodlands. This means that also the conservation of typical ruderal vegetation needs some amount of maintenance in order to preserve it [21, 22, 78].

Management directions are offered by [55] – but even though the conservation and fostering of ruderal vegetation demands some care, the amount of manpower and financial expense is lower by far when compared to traditional green spaces because the initial measures of their establishment can be omitted. Also the further expenses for watering, fertilizing and the replanting of dead plants are spared.

Besides the consideration of cost-saving, one should not forget that green spaces have to be attractive to the public. Some authors (see Discussion) have shown that species-richness and the existence of bright-coloured flowers induces attractiveness. “Beauty” is highly relevant for all people in influencing their attitude towards wilderness areas [2]. This paper will show that a particular group of ruderal communities, the associations of the *Dauco-Melilotion*, is both species-rich and contains many beautiful eye-catchers. Additionally, all other criteria mentioned in literature as important for the acceptance of green spaces will be discussed

Methods

The Central European literature on spontaneous urban vegetation was screened for relevés of plant communities of the alliance *Dauco-Melilotion* (*Onopordetalia*, *Artemisietea*). We considered only well-defined communities widespread in most urban-industrial areas of Central Europe:

- *Artemisio-Tanacetetum vulgaris* Siss. 1950 nom. inv.
- *Echio-Melilotetum* R. Tx. 1947

- *Echio-Verbascetum* Siss. 1950 (for differences between *Echio-Melilotetum* and *Echio-Verbascetum* see [81])
- *Berteroëtum incanae* Siss. et Tidemann in Siss. 1950
- *Dauco-Picridetum hieracioides* (Farber 1933) Goers 1966
- *Resedo-Carduetum nutantis* Siss. 1950.

Characteristic habitats of all these communities are urban-industrial fallow areas [80]. Particularly the *Artemisio-Tanacetetum* and the *Echio-Melilotetum* also grow on unused open spaces within residential areas, with fragments even around street trees [77] and on unpaved banquettes in the inner city [41].

The two most important character species of *Echio-Verbascetum*, *Verbascum thapsus* and *Verbascum densiflorum*, only rarely occur in the same plot. Therefore, we differentiated between a *Verbascum thapsus*-facies and a *V. densiflorum*-facies of this community. As we wanted to present the typical (average) colour spectrum of each community, we considered all character species and additionally those species which, when considering all communities, had a sum of constancy classes >VI (constancy classes are explained in Table 1).

From the constancy of the species we calculated the constancy of flower colours. Some species, in particular those that are wind- or self-pollinated, have inconspicuous flowers. We summarized this group, mainly consisting of grasses (Poaceae), *Artemisia vulgaris* and *Erigeron canadensis*, as “grey”. The colour spectra presented in this paper show the percentage of each colour in relation to the constancy sum of all colours represented in the upper layer of the particular community. Ground creepers, rosette species of low size and tiny therophytes were not considered, because they do not contribute to the floral display of these communities. As not all species are in flower at the same time, we also present the flower phenology of each community. The flower period of the species was taken from [65].

Results

The upper layer (height >50 cm) of the communities of *Dauco-Melilotion* (Table 1) contains the following constantly occurring species that have comparatively large (>1 cm) coloured flowers or smaller flowers that form large flower stands (in alphabetical order): *Carduus acanthoides* and *C. nutans* (pink), *Cichorium intybus* (blue), *Cirsium arvense* and *C. vulgare* (pink), *Daucus carota* (white), *Echium vulgare* (blue), *Hypericum perforatum* (yellow), *Linaria vulgaris* (yellow), *Melilotus albus* (white), *M. officinalis* (yellow), *Oenothera* spp. (yellow), *Pastinaca sativa* (yellowish-green), *Picris hieracioides* (yellow), *Tanacetum vulgare* (yellow), *Reseda lutea* (white), *R. luteola* (yellowish-green), *Silene latifolia* subsp. *alba* (white), *Verbascum densiflorum* and *V. thapsus* (yellow).

Table 1: Species composition of *Dauco-Melilotion* communities growing in urban areas in Central Europe

Association	1	2	3	4	5	6	7
Number of relevés	222	226	20	18	72	57	15
<i>Character species of the communities</i>							
<i>Tanacetum vulgare</i>	V	III	II	II	III	II	II
<i>Melilotus albus</i>	I	V	II			II	
<i>Melilotus officinalis</i>		IV	II	I	I	II	
<i>Verbascum densiflorum</i>		I	III	V			
<i>Verbascum thapsus</i>		I	V	I	I	II	
<i>Berteroa incana</i>		II			V		I
<i>Picris hieracioides</i>			II			V	
<i>Carduus nutans</i>							V
<i>Character species of the alliance</i>							
<i>Daucus carota</i>	III	IV	V	IV	III	V	II
<i>Hypericum perforatum</i>	II	II	V	V	II	III	II

<i>Carduus acanthoides</i>	I	I	I	II	I	II	I
<i>Linaria vulgaris</i>	I	I	II	II	I	II	I
<i>Oenothera biennis</i> agg.	I	III	IV	II	II	II	III
<i>Echium vulgare</i>		II	III	II	II	I	II
<i>Reseda lutea</i>		II	II	II	I	I	III
<i>Reseda luteola</i>			I	II	I	I	II
<i>Cichorium intybus</i>	I	I			I	II	
<i>Pastinaca sativa</i>	II	I				I	II
Character species of order and class							
<i>Artemisia vulgaris</i>	V	V	IV	III	IV	IV	V
<i>Solidago canadensis</i>	III	II	II	II	I	I	
<i>Urtica dioica</i>	II	I	III	II		II	I
<i>Silene alba</i>	II	II	II	III	II		II
<i>Cirsium vulgare</i>	II	II	III	II		II	
<i>Convolvulus arvensis</i>	I	II				I	II
<i>Solidago gigantea</i>	I		III				II
Companions, upper stratum							
<i>Dactylis glomerata</i>	IV	III	II	I	III	IV	II
<i>Cirsium arvense</i>	IV	III	IV	II	II	III	II
<i>Elymus repens</i>	III	II	II	I	III	III	II
<i>Erigeron canadensis</i>	II	II	V	III	III	II	III
<i>Tripleurospermum perforatum</i>	I	III	IV	II	II	II	II
<i>Poa pratensis</i> agg.	II	II	I	I	I	II	II
<i>Arrhenatherum elatius</i>	III	II	II	I	II	III	II
<i>Plantago major</i>	I	II	I		I	II	I
<i>Achillea millefolium</i> agg.	III	III		II	III	IV	III
<i>Festuca rubra</i>	II	II		I	I		I
<i>Holcus lanatus</i>	I	II	III	II			I
<i>Poa palustris</i>	I	II	II	I			I
<i>Erigeron annuus</i>	I		II	II	I	I	
<i>Rumex crispus</i>	II		II	I		I	I
Companions, ground stratum							
<i>Medicago lupulina</i>	II	IV	II	II	II	III	I
<i>Plantago lanceolata</i>	III	II	I	II	III	III	III
<i>Taraxacum officinale</i> agg.	III	III	II	II	II	II	II
<i>Poa compressa</i>	I	III	III	II	I	II	II
<i>Trifolium repens</i>	II	II	I	II	II	I	I
<i>Lolium perenne</i>	II	II	I		II	I	II
<i>Arenaria serpyllifolia</i> agg.		II	V		II	I	III
<i>Diplotaxis tenuifolia</i>			II	II	II	I	II
<i>Cerastium holosteoides</i>		I	II	II			II

Footnote: *shortened synoptic table: all character species of the associations and all other species that have a constancy sum of at least VI (e.g. constancy I in six associations, or I in four or five associations and additionally II in another association, or I in three associations and additionally III in another association, or II in three associations etc.) are listed; constancy classes + und r were not considered; constancy classes: V: > 80%; IV: > 60% - 80%; III: > 40% - 60%; II > 20% - 40%; I > 10% - 20%.

Species of the ground layer were not considered for the colour spectra (Fig 1 and 2).

Column 1: *Artemisio-Tanacetum*: Dannenberg (1995): 33 relevés; Dettmar (1992): 12 relevés; Fijałkowski (1967): 8; Frost (1985): 23; Gödde (1986): 30; Hetzel (1988): 11; Hetzel & Ullman (1981): 12; Kienast (1978): 34; Nežadal & Heider (1994): 37; Pyšek & Pyšek (1988): 2; Rostanski & Gutte (1971): 4; Springer (1985): 5; Wittig et al. (1999): 8; Wollert (1991): 2.

Column 2: *Echio-Melilotum*: Brandes (1977): 4; Brandes (1980): 13; Dannenberg (1995): 44; Dettmar (1986): 13; Fijałkowski (1967): 15; Frost (1985): 10; Gödde (1986): 29; Hetzel & Ullman (1981): 4; Kienast (1978): 10; Nežadal & Heider (1994): 10; Olsson (1978): 4; Pyšek (1991): 4; Rebele (1986): 19; Reidl (1989): 13; Rostanski & Gutte (1971): 2; Springer (1985): 6; Wittig et al. (1999): 13; Wollert (1991): 13.

Column 3: *Verbascum thapsus*-community ("Echio-Verbascetum"): Frost (1985): 1; Gödde (1986): 17; Griese (1999): 1; Reidl (1995): 2; Sauerwein (1988): 1; Springer (1985): 3; Wittig (n.p.; Ruhr district, 2000): 3; Wittig et al. (1999): 5.

Column 4: *Verbascum densiflorum*-community ("Echio-Verbascetum"): Gödde (1986): 9; Griese (1999): 1; Springer (1985): 3; Wittig et al. (1999): 5.

Column 5: *Berteroetum incanae*: Bornkamm (1973): 1; Brandes (1977): 6; Brederock (n.p. Frankfurt/M, railway stations, 1993): 6; Gödde (1986): 12; Hetzel (1988): 2; Hetzel & Ullmann (1981): 1; Lotz (n.p.; Frankfurt/Main, east port, 1993): 5; Nezdal & Heider (1994): 23; Rebele (1986): 3; Springer (1985): 8; Wittig et al. (1999): 5.
 Column 6: *Dauco-Picridetum*: Brandes (1977): 8; Brandes (1989): 2; Brederock (n.p., Frankfurt/M, railway stations, 1993): 5; Dannenberg (1995): 3; Gödde (1986): 8; Hetzel & Ullmann (1981): 16; Kopecky (1982): 3; Lotz (1993): 3; Reidl (1989: see above): 1; Springer (1985): 2; Wittig et al. (1999): 6.
 Column 7: *Resedo-Carduetum nutantis*: Brandes (1989): 1; Dannenberg (1995): 5; Dettmar (1992): 10; Gödde (1986): 6; Hetzel (1988): 1; Sissingh (1950): 4.

Fig. 1 shows that all the above-mentioned colours are present in each of the seven communities analyzed. Uncoloured flowers (“grey”) never represent more than 30% of the community, in three cases even less than 25%. Disregarding the uncoloured flowers, yellow and white are most important in six of the seven communities, and pink always occupies the third position. Taking the lowest percentage of uncoloured flowers for ranking the colourfulness, the *Verbascum densiflorum* facies of *Echio-Verbascetum* takes first position. *Resedo-Carduetum* shows the most equal distribution of all colours, but has the second highest ratio of “grey”.

The flowering of the *Dauco-Melilotum* communities (Fig. 2) starts in May with six species and three colours (yellow, white and pink). In June, the colour optimum is reached (all colours present), but the number of species with coloured flowers increases up to August (June 13, July 23, August 24). For September a slight, and for October a strong, decrease in the number of coloured flowers has to be admitted, but, up until the end of the vegetation period, 11 species are in flower and all colours are represented.

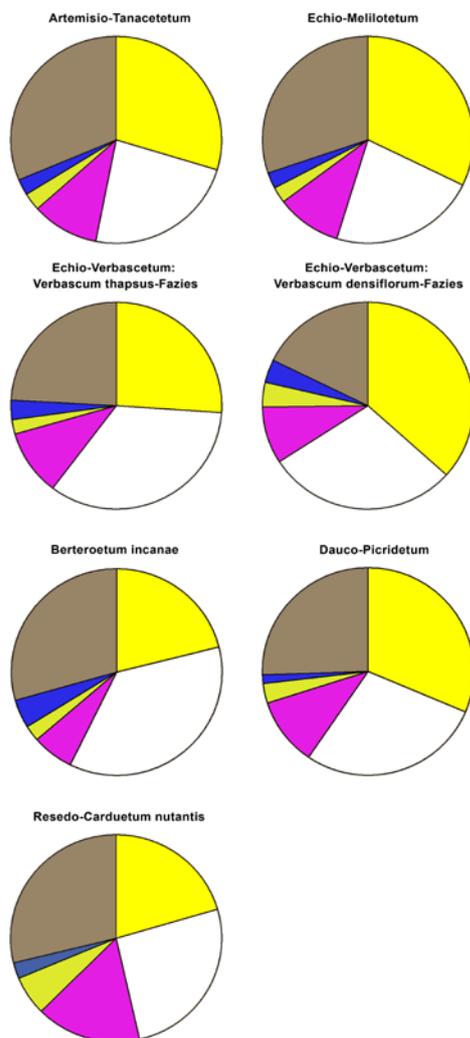


Fig. 1: Colour-spectra of Dauco-Melilotum communities

	2	3	4	5	6	7	8	9	10	11	12
<i>Silene alba</i>											
<i>Melilotus officinalis</i>				Yellow	Yellow	Yellow	Yellow	Yellow			
<i>Melilotus albus</i>											
<i>Daucus carota</i>											
<i>Convolvulus arvensis</i>											
<i>Carduus acanthoides</i>				Pink	Pink	Pink	Pink	Pink			
<i>Oenothera biennis</i> agg.					Yellow	Yellow	Yellow	Yellow			
<i>Hypericum perforatum</i>											
<i>Reseda luteola</i>											
<i>Cirsium arvense</i>						Pink	Pink	Pink			
<i>Echium vulgare</i>						Blue	Blue	Blue	Blue		
<i>Berteroetum incanae</i>											
<i>Tripleurospermum perforatum</i>											
<i>Erigeron annuus</i>											
<i>Pastinaca sativa</i>							Yellow	Yellow			
<i>Carduus nutans</i>							Pink	Pink			
<i>Reseda lutea</i>											
<i>Achillea millefolium</i>											
<i>Cichorium intybus</i>							Blue	Blue			
<i>Tanacetum vulgare</i>							Yellow	Yellow			
<i>Verbascum thapsus</i>							Yellow	Yellow			
<i>Verbascum densiflorum</i>							Yellow	Yellow			
<i>Pieris hieracioides</i>											
<i>Solidago canadensis</i>							Yellow	Yellow			
<i>Cirsium vulgare</i>									Pink	Pink	
<i>Solidago gigantea</i>									Yellow	Yellow	

Fig. 2: Colour-phenology of Dauco-Melilotum communities

Discussion

Public awareness of and interest in biodiversity is particularly influenced by people's experience of nature in their everyday life [64]. When asked for their spontaneous associations considering the term „biodiversity“, the majority of those interviewed named phenomena not belonging to their everyday life, e.g. rainforests and oceans [59]. “If there is to be broad-based public support for biodiversity conservation, the places where people live and work should be designed so as to provide opportunities for meaningful interventions with the natural world” [45, p. 430; see also 46]. For inhabitants of metropolises and large urban agglomerations, natural and rural environments are situated far away and therefore not part of daily life. Particularly for children, there is an urgent need for chances to experience “nature”, “wilderness”, “biodiversity”, or however we might call spontaneously developed urban vegetation and the adherent biocoenoses, within their immediate surroundings. “Here, nature is 'up close and personal' rather than distant or remote” [47, p.23]. This close contact provides many opportunities for people to learn about, and appreciate, wildlife [71]. The critical age of influence seems to be lower than 12 years [4]. According to a survey of 860 households in Halle an der Saale (Germany), people agree that "children need a place within the urban housing environment where they can play in the open air and can experience nature” [51]. Traditional urban parks are not ideally suited for allowing children to have “positive adventures” in the natural environment, because the traditional park vegetation is not tolerant of disturbance [22]. In contrast, „wilderness in the city“ offers children a positive experience of nature [27]. Spontaneous ruderal communities such as *Echio-Verbascetum* and *Artemisio-Tanacetetum* have been rated as a vegetation type well-suited to nature experience areas [43].

As already mentioned in the Introduction, even in those countries that have some experience in “going spontaneous” [29], i.e. in declaring urban ‘brown fields’ as part of the urban green space, ruderal vegetation is not generally appreciated. Still “the aesthetic experience delivered by derelict sites is predominantly governed by disharmony, disarray and chaos, instead of harmony, order and unity” [50, p. 271]. But "ecology will not become a driving force in the development of society and our cities in particular without aesthetics. Ecological aesthetics must attack the contradictions between urban landscapes and other projects of social, political and economic life" [30, p. 306]. Therefore, in order to make spontaneous vegetation more attractive and alternative to ornamental plantings in the city” [40], we have to ask what is rated as “aesthetic” and „attractive“ or, in more simple terms, what is liked and what is disliked.

According to [72], preference or liking is comparatively high, if the following conditions are fulfilled:

- complexity (moderate to high number of independently perceived elements in the scene);
- existence of a focal point, and other order or pattern also present;
- ground surface judged by the observer as favourable for physical activities;
- a deflected or curving sightline;
- perceived threat is negligible, see also [9].

Generally, the three latter criteria are met by sites covered by the communities treated above, or, if not, can easily be achieved. As such areas are frequently used for walking, jogging or cycling [33], they are obviously perceived as favourable for physical activities. A deflected or curving sightline sometimes is present, resulting from the history of the area or can be obtained by designing curved footpaths and allowing some (not to much: see below) trees and shrubs to grow along these paths. In open landscapes, as formed by the low vegetation of the *Dauco-Melilotion* communities, perceived threat is negligible. But a high density of shrubs has strong negative effects on acceptance. Low shrub density and grassy or herbaceous ground cover

provoke strong positive effects [72], and this is exactly the characteristic picture of areas covered by Dauco-Melilotion.

The existence of some trees and shrubs will create a higher complexity and represent focal points, i.e. it will contribute to meet the first two criteria mentioned above for liking. But according to [74] only such members of a set of objects that are sufficiently distinct from their surroundings are visually labelled and therefore consciously perceived. Benkowitz & Köhler [3] show the importance of a species-rich vegetation: from an aesthetic point of view most children preferred a species-rich vegetation compared to a species-poor vegetation. "The reason for their choice was, in general, the presence of particular 'eye-catching' plants."

This is a problem for many ruderal communities, because their dominant species are wind-pollinated [80, 36] and for this reason have small, inconspicuous flowers, so that the total area covered by them is more or less regarded as boring. But our results show that this is not true for the Dauco-Melilotion communities. On the contrary, this urban-industrial vegetation is characterized and dominated by many eye-catchers (Fig.3–5).

According to our own observations, Artemiso-Tanacetetum represents a rather stable stage that can persist from five to ten years. All other Dauco-Melilotion communities will be replaced by less attractive vegetation types (*Rubus armeniacus* community, *Calamagrostis epigejos* stands) after a shorter time, in some cases with an interregnum of Artemiso-Tanacetetum, ultimately leading to urban forest, e.g. Epilobio-Salicetum capreae, *Robinia pseudacacia* forest, *Ailanthus altissima* forest. To guarantee the existence of the Dauco-Melilotion communities, succession has to be suppressed or periodically reset. This is not necessary over the whole area, because in some small and scattered plots undisturbed succession can be allowed. Activities of the users will prevent some other parts of the area from succession. A mosaic of different successional stages will also favour a high diversity of phytophagous insects [70].

Planners obviously keep it self-evident that derelict urban land has to be converted into accepted new landscapes [37] and that this includes vegetation management [24]. In the Ruhr district, pieces of art have been placed in the fallow areas in order to serve as focal points and to demonstrate that these areas are not neglected [13]. Therefore, it should be no problem to convince urban stakeholders and decision makers into the necessity of such management measures. Remembering the criteria of [72], not all the entire remaining area has to be maintained as covered by colourful flowering communities. Some randomly distributed plots would be enough to function as focal points. On the other hand, many investigations have proved that just the early successional stages are important habitats of rare and threatened species, [e.g.17, 32]. Therefore, from the point of biodiversity, the maintenance of larger areas of these communities is appreciated.

The Dauco-Melilotion communities present their rich colour optimum from June to September, i.e. a period of four months. This is much longer than what can be stated for most flowerbeds in traditional parks where the set of species has to be replaced three or four times per vegetation period in order to maintain colour. Dauco-Melilotion communities show no colour only at the beginning of the vegetation period. However, if we accept that there has to be some management, this problem can easily be solved by planting some spring-flowering geophytes. In particular, *Chionodoxa* spp. (blue), *Eranthis hyemalis* (yellow), *Galanthus nivalis* (white), *Muscari armeniacum* (blue) and *Scilla biflora* (blue) that are frequently cultivated in gardens have already successfully escaped in most settlements [82], but, having no mechanisms for long-distance dispersal, are at present generally restricted to the immediate environs of gardens. Planting them in spontaneous ruderal vegetation would only represent anticipation of their spontaneous spreading.



Fig. 3: Echio-Melilotetum with high proportion of *Oenothera* spp. (Dortmund, area of the former coking plant Hansa; photo: Wittig 7/2001)

What is rated as “aesthetic” depends upon user-group differences in terms of background variables such as familiarity and education level [73]. The vegetation of urban green spaces is highly “influenced by the aesthetic and practical needs of the local stakeholders and actors and those having differing interest in different places. Public decision-makers and private individuals belonging to the middle and upper classes have an interest in representative, ornamental and open plantings within the city to demonstrate prosperity, order and security” [62, p. 286]. Therefore, we have to understand that, in spite of some positive examples, where the users have learnt that these areas can meet not only all of the functions expected for the traditional urban green, but even some more that are not allowed in traditional parks, like making a fire, mountain-biking, picking flowers [e.g. 33], in general “the process of changing people's perception on native 'weeds' will take time and a lot of education” [69, p. 306]. Information and education campaigns [e.g. 83], are needed to convince more people of the beauty and importance of spontaneous ruderal vegetation.

As in large parts of Central Europe summers will become drier and hotter and many of the typical ornamental plants, as well as most lawns, already need irrigation under present climate conditions, park planners and managers have to reconsider their concepts with regard to the choice of adapted species. Spontaneous ruderal vegetation is adapted in any case, because its composition is the result of the habitat conditions. Giving more space to this vegetation type would save costs and water and thus represent a substantial step towards a sustainable city.



Fig. 4: *Echio-Verbascetum*, *Verbascum densiflorum* facies (Oberhausen, area of the former: coal mine *Vondern*; photo: Wittig 7/2000)

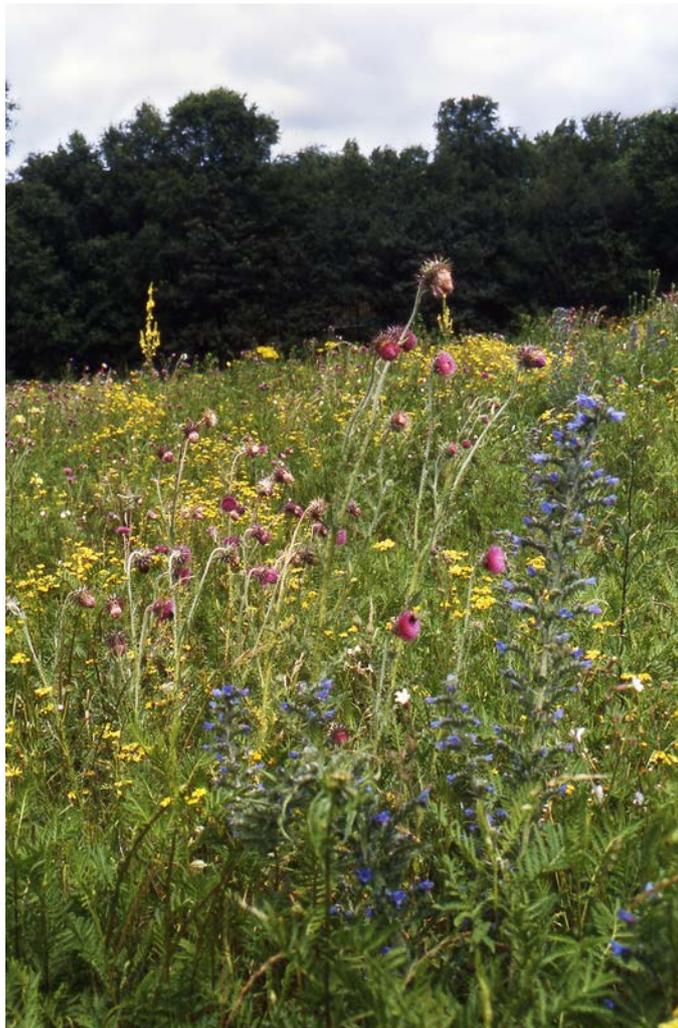


Fig. 5: *Resedo-Carduetum nutantis* (Oberhausen, area of the former: coal mine *Vondern*; photo: Wittig 7/2000)

REFERENCES

1. Auhagen, A., Sukopp, H., 1983. Ziel, Begründungen und Methoden des Naturschutzes im Rahmen der Stadtentwicklungspolitik von Berlin (Aim, reasons and methods of nature protection in the frame of the urban development policy of Berlin). *Natur Landschaft* 58, 9-15 (in German).
2. Bauer, N., 2005. Attitudes toward wilderness and public demands on wilderness areas, in: Kowarik, I., Körner, S. (Eds.), *Wild urban woodlands*. Springer, Berlin, Heidelberg, pp. 47-66.
3. Benkowitz, D., Köhler, K., 2010. Perception of biodiversity – the impact of school gardening, in: Müller, N., Werner, P., Kelcey, J.G. (Eds.), *Urban biodiversity and design*. Wiley-Blackwell, Chichester, pp. 425-440.
4. Bird, W., 2007. Natural thinking. Investigating the links between the natural environment, biodiversity and mental health. A report for the Royal Society of Birds. Available from website: http://www.rspb.org.uk/Images/naturalthinking_tcm9-161856.pdf
5. Brandes, D., 1977. Die Onopordion-Gesellschaften der Umgebung Braunschweigs (The Onopodion communities of the environs of Braunschweig). *Mitt. flor.-soz. Arbeitsgem.* N.F. 19/29, 103-113 (in German).
6. Brandes, D., 1980. Die Ruderalvegetation des Kreises Kelheim Teil 1 (The ruderal vegetation of the district of Kelheim, part 1). *Hoppea* 39, 203-234 (in German).
7. Brandes, D., 1982. Die Gefährdung der städtischen Vegetation: Das Beispiel Braunschweig (Endangerment of the urban vegetation: The example of Braunschweig). *Mitt. TU Carolo-Wilhelmina zu Braunschweig* 17(1), 63-68 (in German).
8. Breuste, J.H., 2004. Decision making, planning and design for the conservation of indigenous vegetation within urban development. *Landscape Urban Planning* 68, 439-452.
9. Burgess, J., Harrison, C.M., Limb, M., 1988. People, parks and urban green: a study of popular meanings and values for open spaces in the city. *Urban Studies* 25, 455-473.
10. Dannenberg, A., 1995. Die Ruderalvegetation der Klasse Artemisietea vulgaris in Schleswig-Holstein (the ruderal vegetation of the class Artemisietea in Schleswig-Holstein). *Mitt. Arb.gem. Geobot. Schleswig-Holstein u. Hamburg* 49, 142 p. (in German).
11. Dettmar, J., 1986. Spontane Vegetation auf Industrieflächen in Lübeck (Spontaneous vegetation on industrial areas in Lübeck). *Kieler Notizen* 18, 113-148 (in German).
12. Dettmar, J., 1992. Industrietypische Flora und Vegetation im Ruhrgebiet (The typical flora and vegetation in the Ruhr industrial area). *Diss. Bot.* 191, 397 p.
13. Dettmar, J., Ganser, K., 1999. *IndustrieNatur – Ökologie und Gartenkunst im Emscher Park*. Ulmer, Stuttgart (in German).
14. Fijałkowski, D., 1967. Zbiorowiska roślin synantropijnych miasta Lublina (Communities of synanthropic plants in the town area of Lublin). *Ann. Bot. Univ. Mariae Curie-Skłodowska, Sectio C*, 12, 195-233 (Polish with Engl. abstract).
15. Frost, D., 1985. Untersuchungen zur spontanen Vegetation im Stadtgebiet von Regensburg (Investigations on the spontaneous vegetation in the urban area of Regensburg). *Hoppea* 44, 5-83 (in German).
16. Galindo, M.P.G., Rodriguez, J.A.C., 2000. Environmental aesthetics and psychological wellbeing: relationships between preference judgments for urban landscapes and other relevant affective responses. *Psychology Spain* 4, 13-27.
17. Gibson, C.W.D., 1998. Brownfield: red data. *The values artificial habitats have for uncommon invertebrates*. English Nature Res. Rep. 273.
18. Gödde, M., 1986. *Vergleichende Untersuchungen der Ruderalvegetation der Großstädte Düsseldorf, Essen und Münster* (Comparative studies of the ruderal vegetation of the megalopolises Düsseldorf, Essen and Münster). Diss. Math.-Naturwiss. Fak. Univ. Düsseldorf (in German).
19. Griese, D., 1999. Flora und Vegetation einer neuen Stadt am Beispiel von Wolfsburg (Flora and vegetation of a new city illustrated by the example of Wolfsburg). *Braunschw. Geobot. Arb.* 7, 235 p. (in German).
20. Hard, G., 1984. Spontane und angebaute Vegetation an der Peripherie der Stadt (Spontaneous and cultivated vegetation in the outskirts of a city). *Schr.r. Fachber. Stadtplanung Landschaftsplanung GHS Kassel* 8, 77-113 (in German).
21. Hard, G., 1997. Spontane Vegetation und Naturschutz in der Stadt (Spontaneous vegetation and nature protection in the city). *Geograph. Rundschau* 10/1997, 562-568.
22. Hard, G., 1998. *Ruderalvegetation. Ökologie & Ethnökologie, Ästhetik & "Schutz"* (Ruderal vegetation. Ecology & ethno-ecology, aesthetics and "conservation"). Notizbuch 49 Kasseler Schule, 396 p (in German).
23. Hard, G., Pirner, J., 1988. *Die Lesbarkeit eines Freiraumes* (Readability of an open space). *Garten + Landschaft* 1/1988, 24-30 (in German).
24. Henne, S.K., 2007. Vegetationsmanagement als Methode der Landschaftsarchitektur für unentdeckte Freiräume (Vegetation management as a method of landscape architecture for undiscovered open spaces). *CONTUREC* 2, 107-115 (in German).

25. Hetzel, G., 1988. Ruderalvegetation im Stadtgebiet von Aschaffenburg (Ruderal vegetation in the city of Augsburg). *Tuexenia* 8, 211-238 (in German).
26. Hetzel, G., Ullmann, I., 1981. *Wildkräuter im Stadtbild Würzburgs* (Wild herbs in the landscape of Würzburg). Würzburger Universitätsschr. Regionalforsch. 3, 150 p., app. (in German).
27. Heuser, J., 2007. Wildnis für Kinder in der Stadt (Wilderness for children in the city). *CONTUREC* 2, 153-157 (in German).
28. Höhn, U., Jürgens, C., Otto, K.-H., Prey, G., Piniek, S., Schmitt, T., 2007. Industriewälder als Bausteine innovativer Flächenentwicklung in postindustriellen Stadtlandschaften – Ansätze zu einer integrativen wissenschaftlichen Betrachtung am Beispiel des Ruhrgebietes (Industrial woodlands as building blocks for innovative site development in post-industrial urban landscapes – Approches to an integrated scientific analysis using the example of the Ruhr region). *CONTUREC* 2, 53-67 (in German).
29. Ignatieva, M., 2010. Design and future of urban biodiversity, in: Müller, N., Werner, P., Kelcey, J.G., (Eds.), *Urban biodiversity and design*. Wiley-Blackwell, Chichester, pp. 118-144.
30. Ipsen, D., 1998. Ecology as urban culture, in: Breuste, J., Feldmann, H., Uhlmann, O. (Eds.), *Urban Ecology*. Springer, Berlin/Heidelberg/New York etc., pp. 302-306.
31. Jim, C.Y., Chen, W.Y., 2006. Perception and attitude of residents toward urban green spaces in Gunanzhou (China). *Environmental Management* 38, 338-349.
32. Junghans, T., 2007. Urban-industrielle Flächen als "Hotspots" der Blütenpflanzen-Vielfalt am Beispiel der Bahn- und Hafenanlagen von Mannheim (Baden-Württemberg) (Urban-industrial sites as hotspots of flowering plant diversity – Exemplified by harbour and rail facilities in Mannheim, Baden-Württemberg). *CONTUREC* 2, 87-94 (in German).
33. Keil, A., 2005. Use and perception of post-industrial urban landscapes in the Ruhr, in: Kowarik, I., Körner, S., (Eds.), *Wild urban woodlands*. Springer, Berlin, Heidelberg, pp 117-130.
34. Kienast, D., 1978. Die spontane Vegetation der Stadt Kassel in Abhängigkeit von bau- und stadtstrukturellen Quartierstypen (The spontaneous vegetation of the city of Kassel in dependency of city structural area types). *Urbs et regio* 10, 414 p., app. (in German).
35. Kinzig, A.P., Warren, P., Martin, C., Hope, D., Katti, M., 2005. The effects of human socioeconomic status and cultural characteristics on urban patterns of biodiversity. *Ecology Society* 10: p 23, [http://zimmer.csufresno.edu/~mkatti/mkatti/Publications_files/Kinzig_et_al_2005.pdf].
36. Knapp, S., Kühn, I., Wittig, R., Ozinga, W.A., Poschlod, P. & Klotz, S., 2008. Urbanization causes shifts in species' trait state frequencies. *Preslia* 80, 375–388.
37. Körner, S., 2007. Natur in der urbanen Landschaft (Nature in the urban landscape). *CONTUREC* 2, 5-13 (in German).
38. Kopecký, K., 1982. Die Ruderalpflanzengesellschaften im südwestlichen Teil von Praha (4) (The ruderal plant communities in the south-western part of Prague (4)). *Preslia* 54, 123-139 (in Czech).
39. Kowarik, I., Langer, A., 1994. Vegetation einer Berliner Eisenbahnfläche (Schöneberger Südgelände) im vierten Jahrzehnt der Sukzession (The vegetation of a Berlin railway area (Schöneberger Südgelände) in the 4th decade of succession). *Verh. Bot. Ver. Berlin Brandenburg* 127, 5-43 (in German).
40. Kuhn, N., 2006. Intentions for the unintentional spontaneous vegetation as the basis for innovative planting design in urban areas. *Journal of Landscape Architecture* 2, 46-53.
41. Langer, A., 1994. *Flora und Vegetation städtischer Straßen am Beispiel Berlins* (Flora and vegetation of city streets using the example of Berlin). Landesentwickl. Umweltforsch., TU Berlin, Sonderh. 10 (in German).
42. Mathey, J., Rink, D., 2010. Urban Wastelands – a chance for biodiversity in cities? Ecological aspects, social perceptions and acceptance of wilderness by residents, in: Müller, N., Werner, P., Kelcey, J.G. (Eds.), *Urban biodiversity and design*. Wiley-Blackwell, Chichester, pp. 406-424.
43. Mathey, J., Kochan, B., Stutzriemer, S., 2003. Städtische Brachflächen – ökologische Aspekte in der Planungspraxis (City brownfields – ecological aspects of urban planning praxis), in: Arlt, G., Kowarik, I., Mathey, J., Rebele, F., (Eds.), *Urbane Innenentwicklung in Ökologie und Planung*. Inst. Ökol. Raumplanung, Dresden, pp. 75-84 (in German).
44. Maurer, U., Peschel, T., Schmitz, S., 2000. The flora of selected urban land-use types in Berlin and Potsdam with regard to nature conservation in cities. *Landscape Urban Planning* 46, 209-215.
45. Miller, J.R., 2005. Biodiversity conservation and the extinction of experience. *Trends Ecology Evolution* 20, 430-434.
46. Miller, J.R., Hobbs, R.J., 2002. Conservation where people live and work. *Conserv. Biol.* 16, 330-337.
47. Müller, N., Werner, P., 2010. Urban Biodiversity and the case for implementing the convention on biological diversity in towns and cities, in: Müller, N., Werner, P., Kelcey, J.G. (Eds.), *Urban biodiversity and design*. Wiley-Blackwell, Chichester, pp. 3-33.
48. Nezadal, W., Heider, G., 1994. Ruderalpflanzengesellschaften der Stadt Erlangen Teil II: Mehrjährige Ruderalgesellschaften (Artemisietea) (Ruderal plant communities of the city of Erlangen Part II: perennial ruderal communities (Artemisietea). *Hoppea* 55, 193-253 (in German).

49. Olsson, H., 1978. Vegetation of artificial habitats in northern Malmö and environs. *Vegetatio* 36, 65-82.
50. Nohl, W., 1998. Is there such a thing as the esthetics of sustainable cities and what could it be like?, in: Breuste, J., Feldmann, H., Uhlmann, O. (Eds.) *Urban Ecology*. Springer, Berlin/Heidelberg/New York etc., pp. 267-272.
51. Patz, R., Kuhpfahl, I., 1998. Green areas in the city – acceptance of changes. Halle (Saale) as an example, in: Breuste, J., Feldmann, H., Uhlmann, O. (Eds.) *Urban Ecology*. Springer, Berlin/Heidelberg/New York etc., pp. 404-406.
52. Pyšek, P., 1991. Die Siedlungsvegetation des Böhmisches Karsts. 1 (The vegetation of settlements of the Bohemian karst). Syntaxonomie. *Folia Geobot. Phytotax.* 26, 225-261 (German with engl. abstract and summary).
53. Pyšek, P., Pyšek, A., 1988. Die Vegetation der Betriebe der östlichen Teile von Praha 2. Vegetationsverhältnisse (The vegetation of industrial habitats in the eastern part of Praha 2. Vegetation conditions. *Preslia* 60, 339-347 (in Czech with engl summary).
54. Rebele, F., 1986. *Die Ruderalvegetation von Berlin (West) und deren Immissionsbelastung. Landschaftsentwicklung und Umweltforschung* (Ruderal vegetation of Berlin (West) and its immission loads). Schr.r. FB Landschaftsentwicklung der TU Berlin 43, Berlin, 223 pp. (in German).
55. Rebele, F., Dettmar, J., 1996. *Industriebrachen. Ökologie und Management* (Brownfields. Ecology and Management). Ulmer, Stuttgart (in German).
56. Reidl, K., 1989. *Floristische und vegetationskundliche Untersuchungen als Grundlage für den Arten- und Biotopschutz in der Stadt - dargestellt am Beispiel Essen* (Floristic and phytosociological studies as basis for species and biotope conservation – illustrated by the example of Essen). Diss. FB 9 Univ. GHS Essen (in German).
57. Reidl, K., 1995. Flora und Vegetation des ehemaligen Sammelbahnhofes Essen-Frintrop (Flora and vegetation of the former railroad marshalling yard Essen-Frintrop). *Flor. Rundbr.* 29, 68-85 (in German).
58. Reidl, K., Schemel, H.-J., Blinkert, B., 2007. Naturerfahrungsräume – Ein Ansatz zur Naturvermittlung in Stadtgebieten (Places for nature discovery – Ways to provide an experience of nature in urban areas). *CONTUREC* 2, 141-151 94 (in German).
59. Rink, D., 2002. Biodiversität – eine alltägliche Erfahrung? (Biodiversity – a daily experience?) *Denkanstöße* 7(4), 6-72 (in German).
60. Rink, D., Emmerich, R., 2005. Surrogate nature or wilderness? Social perceptions and notions in an urban context, in: Kowarik, I., Körner, S. (Eds.): *Wild urban woodlands*. Springer, Berlin, Heidelberg, pp. 467-80.
61. Rostanski, K., Gutte, P., 1971. Roślinność ruderalna miasta Wrocławia (Ruderal vegetation in Wrocław). *Mater. Zakł. Fitosoc. Stos. Uniw. Warsz.* 27, 167-215 (in Polish).
62. Sattler, D., Schmidt, S., Vinicius da Silva Alves, M., 2010. Analysis of the planted and spontaneous vegetation at selected open spaces in Apipucos district of Recife, Brazil, in: Müller, N., Werner, P., Kelcey, J.G. (Eds.) *Urban biodiversity and design*. Wiley-Blackwell, Chichester, pp 273-290.
63. Sauerwein, B., 1988. Die Pflanzengesellschaften der Henschelhalde in Kassel (Plant communities of the "Henschelhalde" in Kassel). *Philippia* 6, 3-35 (in German).
64. Savard, J.-P., Clergeau, P., Mennechez, G., 2000. Biodiversity concepts and urban ecosystems. *Landscape Urban Planning* 48, 131-142.
65. Senghas, K., Seybold, S., 2003, Schmeil – Fitschen: Flora von Deutschland und angrenzender Länder, 92. Aufl. – Quelle & Meyer, Wiebelsheim, 864 S
66. Sissingh, G., 1950. *Onkruid-Associaties in Nederland* (Weed communities in the Netherlands). Versl. Landbouwk. Onderz. 56, 224 pp., app. (in Dutch).
67. Smardon, R.C., 1988. Perception and Aesthetics of the urban environment: review of the role of vegetation. *Landscape Urban Planning* 15, 85-106.
68. Springer, S., 1985. Spontane Vegetation in München (spontaneous vegetation in Munich). *Ber. Bayer. Bot. Ges.* 56, 103-142 (in German).
69. Stewart, G.H., Ignatieva, M., Meurk, C.D., 2010. Multivariate approaches to the study of urban biodiversity and vegetation: an example from a southern temperate colonial city, Christchurch, New Zealand, in: Müller, N., Werner, P., Kelcey, J.G. (Eds.) *Urban biodiversity and design*. Wiley-Blackwell, Chichester, pp. 291-308.
70. Strauss, B., Biedermann, R., 2006. Urban brownfields as temporary habitats: driving forces for the diversity of phytophagous insects. *Ecography* 29, 928-940.
71. Sundseth, K., Raeymaekers, G., 2006. *Biodiversity and Natura 2000 in urban areas. Nature in cities across Europe: a review of key issues and experience*. Bruxelles Environment-IBGE/Leefmilieu Brussel-BIM, Brussel.
72. Ulrich, R.S., 1986. Human responses to vegetation and landscapes. *Landscape Urban Planning* 13, 29-44.
73. Van den Berg, A.E., Vlek, C.A.J., Coeterier, J.F., 1998. Group differences in the aesthetic evaluation of nature development plans: a multilevel approach. *J. Environm. Psych.* 18(2), 141-157.

74. Wandersee, J.H., Schussler, E.E., 2008. Toward a theory of plant blindness. *Plant Sci. Bull.* 47 [as retrieved from the internet on 28.01.2011]
75. Weiss, J., Burghardt, W., Gausmann, P., Haag, R., Haeupler, H., Hamann, M., Leder, B., Schulte, A., Stempelmann, I., 2005. Nature returns in abandoned industrial land: monitoring succession in urban-industrial woodlands in the German Ruhr, in: Kowarik, I., Körner, S. (Eds.) *Wild urban woodlands*. Springer, Berlin, Heidelberg, pp 143-162.
76. Wittig, R., 1991. *Ökologie der Großstadtflora* (Ecology of the city flora). UTB G. Fischer, Stuttgart, 261 pp.
77. Wittig, R., 1995. Überblick über die Baumscheibenvegetation sechs mitteleuropäischer Städte (Overview of the vegetation of the area around street trees in six Middle European cities). *Schr.r. Vegetationskunde* 27, 231-238.
78. Wittig, R., 1997. Maßnahmen zur Artenanreicherung in Vegetationsbeständen städtischer Grünanlagen (Actions for species enrichment in species inventories of urban green spaces), in: Forschungsges. Landschaftsentwicklung Landschaftsbau e. V. (eds.) *Anlage und Pflege von Grünflächen in der Stadt*, Teil 6: Biotoppflege, Biotopentwicklung, pp 50-58.
79. Wittig, R., (Ed.) 2001. *Nutzbarkeit und Attraktivität von Stadtnatur* (Usability and attractiveness of urban nature). Geobot. Kolloq. 14, 62 pp (in German).
80. Wittig, R., 2002. *Siedlungsvegetation* (Vegetation of settlements). Ulmer, Stuttgart, 252 pp.
81. Wittig, R., 2005. Zur Syntaxonomie von ruderalen Melilotus-, Echium- und Verbascum-Fluren.- *Tuexenia* 25, 195-210.
82. Wittig, R., 2008. Gartenflüchtlinge als neue Mitglieder der Dorfflora in Nordrhein-Westfalen (Garden escapees as new members of the village flora in Northrhine-Westphalia). *Braunschw. Geobot. Arb.* 9, 481-490 (in German).
83. Wittig, R., Krohmer, J., 2008. Banken, Börse... Biodiversität. Eine regionale Bildungskampagne in Frankfurt/Rhein-Main (Banks, stock-markets... biodiversity. A regional education campaign in Frankfurt/Rhein-Main). *Forum Geoökol.* 19(3), 32-36.
84. Wittig, R., Zucchi, H., (Eds.) 1993: Städtische Brachflächen und ihre Bedeutung aus Sicht von Ökologie, Umwelterziehung und Planung (Urban brownfields and their importance from the view point of ecology, environmental education and urban planning). *Geobot. Kolloq.* 9, 79 pp. (in German).
85. Wittig, R., Lenker, K.-H., Tokhtar, V., 1999. Zur Soziologie von Arten der Gattung *Oenothera* L. im Rheintal von Arnheim (NL) bis Mulhouse (F) (Phytosociological relationships of *Oenothera* species in the Rheine Valley between Arnheim (NL) and Mulhouse (F)). *Tuexenia* 19, 447-467 (in German with Engl abstract).
86. Wittig, R., Breuste, J., Finke, L., Kleyer, M., Rebele, F., Reidl, K., Schulte, W., Werner, P., 2008. What should an ideal city look like from an ecological point of view? – Ecological demands on the future city, in: Marzluff, J.M., Shulenberg, E., Endlicher, W., Alberti, M. (Eds.) *Urban Ecology*. Springer, New York, pp. 691-697.
87. Wollert, H., 1991. *Die Ruderalvegetation des Meßtischblattes Teterow* (2241; Mittelmecklenburg) (The ruderal vegetation in the area of the German topographic map Teterow (2241; Mittelmecklenburg)). *Gleditschia* 19, 39-68 (in German).

ESTETICA VEGETAȚIEI SPONTANE URBANO-INDUSTRIALE

(Rezumat)

În orașele aglomerate nu toate terenurile pot fi reciclate și utilizate în scop urbano-industrial. Totuși, datorită problemelor financiare, spațiile deschise rămase nu pot fi toate transformate în parcuri urbane, deoarece astfel de spații verzi necesită multă îngrijire: trebuie să fie proiectate, pregătite (pregătirea solului și a reliefului), plantate și îngrijite (irigații, cosiri, tăieri, replantări). Din aceste motive, în unele regiuni ale Europei, mai ales în zona Ruhr (Germania), terenurile urbano-industriale, acoperite de vegetație spontană au fost declarate oficial spații urbane verzi. Deși argumentele sunt solide, conceptul de vegetație spontană ca și component major al spațiilor verzi urbane nu este de obicei acceptat în Europa centrală și cu atât mai puțin înafara ei. Motivul este acela că vegetația ruderală este considerată de către majoritatea oamenilor ca un indicator de neglijență, abandon și de aceea este văzută ca inestetică și murdară.

Luând în considerare toate releveele disponibile, această lucrare demonstrează că un grup particular de comunități ruderală, asociația *Dauco-Melilotion*, este bogat în specii și conține multe plante frumoase și atractive. Florile incolore („gri”) nu reprezintă niciodată mai mult de 30 % din asociație, de cele mai multe ori chiar mai puțin

de 25 %. Galbenul este cel mai important, în toate cazurile, cu alb ca a doua opțiune și roz-violet pe locul al treilea. Albastrul și galbenul-verzui sunt întotdeauna prezente.

Pe lângă acestea, toate criteriile menționate în literatura de specialitate ca fiind importante pentru acceptarea spațiilor verzi, sunt întrunite de către acest tip de vegetație. Bogăția de culoare optimă este prezentă în perioada iunie-septembrie – o perioadă de patru luni. Această perioadă reprezintă mult mai mult decât se întâmplă în cazul majorității rabatelor cu flori din parcurile tradiționale. De vreme ce în majoritatea zonelor din Europa centrală verile vor deveni tot mai uscate și calde, iar majoritatea plantelor ornamentale tipice, la fel ca și majoritatea pajiștilor necesită irigații chiar și în condițiile climatice actuale, peisagiștii vor trebui să-și reconsidere conceptele cu privire la alegerea speciilor cu adaptare bună. Vegetația ruderală spontană este oricum adaptată, deoarece structura acesteia este rezultatul condițiilor de mediu. Acordarea unui spațiu mai mare acestui tip de vegetație va permite o experiență mai naturală pentru orașeni și mai multe activități de recreere decât verdele tradițional (deoarece este mai tolerantă la perturbări), va contribui la biodiversitate, va reduce costurile și consumul de apă și de aceea reprezintă un pas spre orașul durabil.

Received: 3.11.2011; Accepted: 7.11.2011.